

## Diversity of soil mites (Acari: Mesostigmata: Gamasina) in various deciduous forest ecosystems of Muntenia region (southern Romania)

MINODORA MANU

Institute of Biology, Department of Ecology, Taxonomy and Nature Protection,  
Splaiul Independenței 296, 060031 Bucharest, Romania  
Corresponding author: Minodora Manu, [mindora\\_stanescu@yahoo.com](mailto:mindora_stanescu@yahoo.com)

(Received on 7 February 2011; Accepted on 15 January 2013)

---

**Abstract:** The main task of the research was to investigate the diversity of predatory soil mites (Gamasina) in 8 types of forest ecosystems: alder (Călugăreni, Clinceanca, Azuga Valley, Cumpătu), oak-hornbeam (Balotești, Băneasa), beech (Șotrile), and fir-beech (Lunca Mare) in Muntenia region. Taxonomical classification and statistical methods used in this study show similarities as well as differences between their predatory mite communities. In total, 467 mites of 57 gamasid species were identified, belonging to 28 genera and 12 families. *Veigaia nemorensis*, *Prozercon fimbriatus*, *P. kochii*, and *P. traegardhi* were most abundant. Geographical position, abiotic factors (soil type, slope angle, soil moisture content, pH) and biotic ones (vegetation structure: herbs, shrubs or trees) were related to differences in gamasid species composition between the investigated forest ecosystems.

**Keywords:** Gamasina, diversity, microclimate, mite, structure, similarity

### INTRODUCTION

Natural forests are complex and very stable ecosystems. These characteristics are due to the specific ecological niches of all species, both aboveground and belowground, in the soil food web (MOORE et al., 2005). The soil ecosystem contains many less studied but often abundant groups of mesofauna, such as soil mites and other microarthropods (COLEMAN & WHITMAN 2005). Gamasids, as one group of soil mites, are predators, influencing population growth of other organisms and controlling the abundance of springtails, soil-dwelling mites, larvae and eggs of insects as well as nematodes and enchytraeids. Therefore they have an indirect effect on the structure and function of ecosystems, affecting decomposition of organic matter, nutrient cycling, and formation of mycorrhiza, being an important factor in soil formation and stabilization processes (KOEHLER 1997, 1999; BEDANO & RUF 2010). In soil they find

the most favourable conditions for their development. The structure and dynamics of their populations are usually specific to each type of ecosystems, depending on vegetation structure, soil type, and microclimate.

The main task of this research was to determine if the type of ecosystem in correlation with some abiotic factors influences the structure of soil mite communities in Romanian deciduous forests.

## MATERIAL AND METHODS

### *Study sites*

The study was conducted in 2004-2007 in 8 forest ecosystems: in Azuga Valley (1); Cumpătu (2), Șotriile (3) and Lunca Mare (4) in Doftana Valley; Balotești (5) and Băneasa (6) near Bucharest; and Călugăreni (7) and Clinceanca (8) in Giurgiu district. All of them are natural forests, within Muntenia region of Romania.

The forests of Azuga Valley (site 1, 45°26'56.0"N, 25°36'15.40"E; altitude 1035 m) and Cumpătu (site 2, 45°21'59.6"N, 25°33'02.50"E; altitude 828 m) represent the plant association *Telekio speciosae-Alnetum incanae* Coldea (1986) 1990. They are located in the middle and lower parts of the Azuga Valley, on alluvial soil (pH 7.1-8.5). The herb layer is composed of species characteristic of Carpathian beech forests, like *Pulmonaria rubra*, *Symphytum cordatum*, *Campanula abietina*, *Carduus personata*, *Chaerophyllum hirsutum*, *Viola biflora*, *Geum rivale*, *Heracleum palmatum*, and *Delphinium elatum*. The tree layer is dominated by *Alnus incana* (species common at high altitudes), comprising also *Acer pseudoplatanus*, *Rosa caninna*, *Salix alba*, and *Sambucus nigra* (FALCĂ et al. 2005a). Soil samples were collected there in 2004.

The forest ecosystem of Șotriile (site 3, 45°13'39.67"N, 25°43'48.61"E; altitude 600 m) is a beech forest with low productivity and moder humus of *Calamagrostis-Luzula* type. The habitat is classified as R4106 South-East Carpathian beech and fir forest with *Hieracium rotundatum*. According to Natura 2000, the habitat type is 9110 *Luzulo-Fagetum* beech forest (DONIȚĂ et al. 2005). This ecosystem is situated on a slope of 30°. The soils are acid brown, brown iron-alluvial and podzolic, oligo- and oligo-mesobasic, with moderate to low soil moisture content, moderate to deep clayey-sandy, with small to moderate edaphic volume. Soil fertility is low. Soil samples were collected there in 2006.

The forest of Lunca Mare (site 4, 45°11'27.09"N, 25°44'50.24"E; altitude 490 m) is a beech forest ecosystem of *Epipactis-Cephalanthera* type, with low productivity and mull-moder humus, on calcicolous and eubasic soils. The habitat is classified as R4111 South Carpathian beech forest of *Fagus sylvatica* and *Abies alba* with *Cephalanthera damassonium* (DONIȚĂ et al. 2005). According to Natura 2000, the habitat is 9150 Medio-European limestone beech forest of the alliance *Cephalanthero-Fagion*. It is situated on a strongly fragmented relief with calcareous rocks on the surface, having a slope angle of 40°. The humid calcareous soil is not uniform in structure, varying from calcicolous, clayey to argillaceous or with mull humus. It has good airflow and drainage. Soil samples were collected there in 2005.

Forest ecosystems of Balotești (site 5, 44°42'45.8"N, 26°08'49.9"E; altitude 100 m) and Băneasa (site 6, 44°29'31.6"N, 26°04'46.6"E; altitude 85 m) were characterized by native trees, xero-mesophilous to mesophilous, meso-eutrophic to eutrophic, frequent from forest steppe to beech forests and from oak forest zones (*Carpinus betulus* L., *Quercus cerris* L., *Quercus robur* L., *Robinia pseudoacacia* L., *Tilia cordata* Miller, *Ulmus minor* Miller) and by trees cultivated for forestry and ornamental purposes, thermo-subthermophilous, frequent from forest steppe to oak forest zones, in forests and their edges with shrubs (*Acer tataricum* L., *Ulmus minor* Miller). At the altitude of 100-200 m a.s.l. of the forest steppe and forest areas, the mean temperature was high (10.0–9.5°C) with mean precipitation quite low (700–800 mm, usually 450–750 mm), typical for these forests. The herb layer was dominated by xero-mesophilous to mesophilous perennials, rarely biennials and annuals, frequent from steppe to beech forest and boreal zones, shaded areas, forests, edges of forests, shrublands, ruderal places, and sometimes segetal weeds: *Alliaria petiolata* (Bieb.) Cavara et Grande, *Bromus arvensis* L., *Chenopodium album* L., *Dactylis glomerata* L., *Daucus carota* L., *Lamium amplexicaule* L., *Plantago media* L., *Poa angustifolia* L., *Prunella vulgaris* L., *Rumex crispus* L., *Stellaria media* (L.) Vill., *Taraxacum officinale* Weber ex. Wiggers, *Trifolium repens* L., *Urtica dioica* L. There were also meso-hygrophilous perennials, frequent from steppe to beech forest zones, and meadows (*Agrostis stolonifera* L.). The forests were represented by small patches of diverse plant associations. Distribution of the herb and shrub layers was discontinuous because of human impact, increased especially in Băneasa forests. The soil is sandy (ONETE & PAUCĂ-COMĂNESCU 2008; MANU 2008). Soil samples were collected there in 2007.

The forests of Călugăreni (site 7, 44°11'25.19"N, 25°57'56.45"E; altitude 54 m) and Clinceanca (site 8, 44°08'39.53"N, 26°07'35.45"E; altitude 85 m) are classified as the association *Stellario nemori-Alnetum glutinosae*, specific to large valleys, where water flow is slower. As a consequence of alluvial accumulation, the land is elevated and becomes suitable for meso-hygrophilous meadows. The gley soils are permanently flooded by rivers, sometimes swampy, neutral to slightly basic (pH 7.0-8.2). The tree layer is dominated by *Alnus glutinosa* (species common at low altitudes), comprising also *Salix alba*, *Fraxinus pallisiae*, *F. excelsior*, and *Acer campestre*. In the herb layer, *Brachypodium sylvaticum*, *Aegopodium podagraria*, *Geum urbanum*, *Allium ursinum*, and *Rubus caesius* are accompanied by meso-hygrophilous species of the classes *Galio-Urticetea* and *Molinio-Arrhenatheretea* (FALCĂ et al. 2005b). Soil samples were collected there in 2004.

In an effort to synthesize the presented material, the studied ecosystems were divided into 5 groups: alder forests of high altitudes (sites 1 and 2: Azuga Valley and Cumpătu); alder forests of low altitudes (sites 7 and 8: Călugăreni and Clinceanca); fir-beech forest (site 3: Șotriile); beech forest (site 4: Lunca Mare); and oak-hornbeam forests (sites 5 and 6: Băneasa and Balotești).

### Sampling

From each ecosystem, a total of 15 samples of soil were collected with MacFadyen corer (5 cm in diameter), to a depth of 10 cm. The soil samples were taken 3 times in one year: 5 in spring (May), 5 in summer (July), and 5 in autumn (September) for each ecosystem. The mites were extracted with a modified Berlese-Tullgren funnel, in ethyl alcohol, and the mite samples were clarified in lactic acid. Mesostigmatid mites there identified to species level (GILYAROV & BREGETOVA 1977; HYATT 1980; KARG 1993; MAŠAN 2003; MAŠAN & FENDA 2004; MAŠAN 2007; GWIAZDOWICZ 2007; MAŠAN & HALLIDAY 2010). In total, 120 soil samples were analysed, with 57 gamasid species and 467 individuals.

To measure the moisture content of the soil, 18 samples per year (6 per season) were collected from each ecosystem. The pH was measured with a C532 Jasco Consort-pH-meter. The moisture content of soil and pH for each area are represented by average annual values with standard deviation (Table 1).

### Mite community analysis

After taxonomic identification, the number of individuals was the basis for calculation of density ( $m^{-2}$ ), dominance index  $D$  (%), constancy index  $C$  (%), and Jaccard similarity index  $J$ .

Density ( $m^{-2}$ ) was calculated using the formula ( $\sum$  no. of individuals/no. of samples) \*  $1m^2$ /surface area of the soil core (BOTNARIUC & VĂDINEANU 1982). The surface area of the soil core was  $19.63\text{ cm}^2$ .

The results were analysed with the aid of BioDiversityPro 2.0 program, to calculate the Jaccard index  $J$  for mite communities from the 8 studied forests.

$$J = c / (a + b - c),$$

where:  $a$  = number of species in ecosystem A;  $b$  = number of species in ecosystem B;  $c$  = number of species common to ecosystems A and B.

The dominance index (%) was calculated using the formula:

$$D = 100\% * n/N,$$

where:  $n$  = number of individuals of one species in all samples;  $N$  = total number of individuals of all species in all samples. Dominance classes for the identified gamasid mites were: eudominants with  $D > 10.0\%$  (D5); dominants with  $D$  of 5.1–10.0% (D4); subdominants with  $D$  of 2.1–5.0% (D3); recedents with  $D$  of 1.1–2.0% (D2), and subrecedents with  $D < 1.1\%$  (D1) (ENGELMANN 1978).

The constancy index (%) was obtained using the formula:

$$C = 100\% * p_A/P,$$

where:  $p_A$  = number of samples with species A;  $P$  = total number of samples. The mite species were classified in 4 constancy classes: euconstant species with  $C$  of 75.1–100% (C4); constant species with  $C$  of 50.1–75% (C3); accessory species with  $C$  of 25.1–50% (C2); and accidental species with  $C$  of 1–25% (C1) (SELVIN & VACCA 2004).

## RESULTS

Mean soil moisture content was the highest (43.75%) at Balotești and the lowest (14.73%) at Șotrile and Clinceanca (15.27%). In other forests, soil moisture content was also low (means between 28.62% and 22.00%). The most acidic soils were identified at Șotrile, Lunca Mare, Clinceanca, and Călugăreni, while basic ones at Cumpătu and Băneasa (Table 1). The recorded values of soil moisture content are not close to the optimum of 60%, but some species are capable to adapt to the dryness, so species richness and diversity also recover rapidly (METZ 1971; SALMANE 2000; MANU & HONCIUC 2010b).

Table 1. Annual averages of soil moisture content and pH in the studied forest ecosystems ( $\pm$  standard deviation)

Factor	Forest sites							
	1	2	3	4	5	6	7	8
Moisture (%)	28.30 $\pm 7.30$	24.93 $\pm 6.67$	14.73 $\pm 6.33$	28.17 $\pm 9.86$	43.75 $\pm 3.77$	28.62 $\pm 4.50$	22.00 $\pm 5.57$	15.27 $\pm 6.12$
pH	7.10 $\pm 0.12$	8.50 $\pm 0.09$	4.89 $\pm 0.28$	5.52 $\pm 0.26$	7.00 $\pm 0.11$	8.20 $\pm 0.08$	5.87 $\pm 0.17$	5.07 $\pm 0.26$

In total, 57 species were identified, belonging to 28 genera and 12 families. Only 1 species occurred in all forests, while 12 were specific to oak-hornbeam forests (Băneasa, Balotești), 5 to beech-fir forest (Lunca Mare), 4 to beech forest (Șotrile), 12 to alder forests of low altitude, 3 to alder forests of high altitude and 20 to alder forests of both types (Călugăreni, Clinceanca, Cumpătu, Azuga Valley). The obtained numbers of gamasid species in the investigated areas are generally similar to those from other temperate forest ecosystems, where this parameter varies from 20 to 98 (KOEHLER 1997, 1999; SKORUPSKI 2001, GWIAZDOWICZ & KLEMT 2004; MORAZA 2006; GULVIK 2007; SKORUPSKI et al. 2008, 2009; SALMANE & BRUMELIS 2010; MANU & HONCIUC 2010a, b).

Considering spatial distribution, numbers of species and individuals were the highest at Călugăreni and Băneasa. Medium values were recorded at Cumpătu, Clinceanca, Lunca Mare, Azuga Valley, and Balotești. The lowest diversity was at Șotrile (possibly due to the steep slope of 40°, which caused instability of the organic layer). Analysing mite density in the investigated areas, the predatory mites had proper conditions for their development at Călugăreni, Băneasa, and Clinceanca, in comparison with the unfavourable conditions at Balotești, Cumpătu, Lunca Mare, Azuga Valley and Șotrile (Table 2).

Table 2. Number of species, individuals, and density of gamasids in the studied forest ecosystems

Site	Forest name	No. of species	No. of individuals	Density (ind./m <sup>2</sup> )
1	Azuga	14	22	747
2	Cumpătu	17	31	1053
3	Șotriile	12	33	1120
4	Lunca Mare	15	62	2105
5	Balotești	14	48	1630
6	Băneasa	21	92	3124
7	Călugăreni	23	93	3158
8	Clinceanca	17	86	2920

Species with the highest density per m<sup>2</sup> were *Leptogamasus parvulus*, *Lysigamasus lapponicus*, *Veigaia nemorensis*, *Pseudolaelaps doderoi*, *Asca aphidoides*, *Rhodacarellus kreuzi*, *Hypoaspis aculeifer*, *Pachylaelaps furcifer*, *Eviphis ostrinus*, *Zercon hungaricus*, *Prozercon kochi*, *P. plumatus*, *P. fimbriatus* and *P. traegardhi* (Table 3).

*Leptogamasus parvulus*, *Lysigamasus lapponicus* and *Veigaia nemorensis* are very mobile predators, able to sustain large populations. *Asca aphidoides* and *Rhodacarellus kreuzi* are small-sized species capable of making horizontal migrations in search of suitable microhabitats in relation to the time of day or prevailing weather (CHRISTIAN 1995; KOEHLER 1999; LINDBERG & BENGTTSSON 2006). *Zercon hungaricus*, *Prozercon kochi*, *P. plumatus*, *P. fimbriatus* and *P. traegardhi* are common eurytopic detritivores. They are most abundant and frequent in leaf litter and soil detritus of deciduous forests, especially with oak (MAŠAN & FENDA 2004). The fact that these species are dominant in all investigated forests means that they apparently do prefer any forest type, being capable to adapt to various environmental conditions.

The high soil moisture content at Călugăreni and Băneasa created favourable habitats for gamasid communities, reflected by the number of identified species and their density per m<sup>2</sup>. Generally, in acid sandy soils the rate of decomposition is low, which could be related to the presence of fungi that generally dominate at low pH. The presence of the fungi from acid soils is associated with abundant fungivores (e.g. nematodes, springtails, enchytraeids, and immature oribatids), which are food sources for gamasids (MARAUN et al. 2003). In spite of the high acidity of soils and of the small edaphic volume at Clinceanca, Lunca Mare, and Șotriile, the number of individuals of gamasids is lower there, due to the decreased moisture content.

The calcareous soil from Cumpătu is not a preferred abiotic condition for gamasid development. In this soil, bacteria are generally more abundant, and their presence is correlated with high soil pH (basic substrate) (KOOIJMAN et al. 2009). Unfortunately, bacteria are not a favourable source of food for predatory mites (WALTER & PROCTOR 1999; MARAUN et al. 2003; BERG & BENGTTSSON 2007).

Two of identified gamasid species are eudominants (*Veigaia nemorensis* and *Prozercon kochi*), one is a dominant (*Prozercon traegardhi*), whereas the remaining 55 species are subdominants, recedents, and subrecedents. Considering the constancy index, 26.31% of species are euconstant; 19.29% are constant, and 55.4% are accessory and accidental species (Table 3).

Table 3. Recorded soil mites (Acari: Mesostigmata: Gamasina) and their population parameters from all investigated forest ecosystems and seasons jointly. SD = standard deviation of number of individuals; *D* = dominance; *C* = constancy

Species	No. of individuals	SD	Density (ind./m <sup>2</sup> )	<i>D</i> (%)	<i>C</i> (%)	Forest type
<b>Epicriidae</b>						
<i>Epicrius mollis</i> (Kramer, 1876)	2	0.89	68	0.43	20	beech
<i>E. tauricus</i> Bregetova, 1977	1	0.45	34	0.22	20	alder (low alt.)
<b>Parasitidae</b>						
<i>Holoparasitus calcaratus</i> (C. L. Koch, 1839)	1	0.45	34	0.22	20	oak-hornbeam
<i>H. cornutus</i> Juvara-Bals & Witaliński, 2000	4	0.84	136	0.87	60	oak-hornbeam
<i>Leptogamasus parvulus</i> (Berlese, 1903)	12	2.51	408	2.60	80	alder (low and high alt.), beech
<i>L. variabilis</i> Juvara-Bals, 1981	1	0.45	34	0.22	20	beech
<i>Leptogamasus</i> sp.	3	0.89	102	0.65	40	oak-hornbeam
<i>Paragamasus similis</i> (Willmann, 1953)	6	0.45	204	1.30	100	alder (low alt.)
<i>Pergamasus quisquiliarum</i> (R & C. Canestrini, 1882)	2	0.55	68	0.43	40	oak-hornbeam
<i>P. laetus</i> Juvara-Bals, 1970	10	1.22	340	2.16	100	alder (low alt.), oak-hornbeam
<i>Pergamasus</i> sp.	7	1.14	238	1.52	80	alder (low and high alt.), beech, oak-hornbeam
<i>Lysigamasus neoruncatellus</i> Schweizer, 1961	1	0.45	34	0.22	20	alder (low alt.)
<i>L. truncus</i> Schweizer, 1961	3	0.89	102	0.65	40	beech, oak-hornbeam
<i>L. lapponicus</i> Tragardh, 1910	7	1.52	238	1.52	60	beech, fir-beech
<i>Lysigamasus</i> sp.	1	0.45	34	0.22	20	oak-hornbeam
<i>L. conus</i> Karg, 1971	1	0.45	34	0.22	20	fir-beech
<i>Parasitus beta</i> Oudemans & Voigts, 1904	2	0.55	68	0.43	40	alder (low and high alt.)
<i>P. furcatus</i> (R & C. Canestrini, 1882)	1	0.45	34	0.22	20	alder (low alt.)
<i>Vulgarogamasus kraepelini</i> (Berlese, 1905)	1	0.45	34	0.22	20	alder (low alt.)

<b>Veigaiidae</b>						
<i>Veigaia nemorensis</i> (C. L. Koch, 1839)	67	4.56	2278	14.50	100	alder (low and high alt.), beech, fir-beech, oak-hornbeam
<i>V. cervus</i> (Kramer, 1876)	1	0.45	34	0.22	20	alder (low alt.)
<i>V. exigua</i> (Berlese, 1917)	8	1.14	272	1.73	80	alder (low alt.), beech, fir-beech, oak-hornbeam
<b>Ascidae</b>						
<i>Cheroseius</i> sp.	1	0.45	34	0.22	20	fir-beech
<i>Arctoseius eremitus</i> (Berlese, 1918)	1	0.45	34	0.22	20	alder (low alt.)
<i>A. cetratus</i> (Sellnick, 1940)	5	1.41	170	1.08	40	alder (low alt.), fir-beech
<i>Leioseius magnanalis</i> (Evans, 1958)	1	0.45	34	0.22	20	alder (low alt.)
<i>Asca aphidoides</i> (Linne, 1758)	4	0.45	136	0.87	80	oak-hornbeam
<i>A. bicornis</i> (Caneastrini & Fanzago, 1887)	3	0.89	102	0.65	40	oak-hornbeam
<i>Protogamasellus singularis</i> (Karg, 1962)	1	0.45	34	0.22	20	oak-hornbeam
<b>Phytoseiidae</b>						
<i>Amblyseius</i> sp.	1	0.45	34	0.22	20	beech
<b>Digamasellidae</b>						
<i>Dendrolaelaps</i> sp.	4	0.84	136	0.87	60	alder (low and high alt.)
<b>Rhodacaridae</b>						
<i>Rhodacarellus kreuzi</i> Karg, 1965	16	4.09	544	3.46	80	alder (low and high alt.)
<i>R. silesiacus</i> Willmann, 1936	5	1.22	170	1.08	60	alder (low and high alt.), oak-hornbeam
<b>Macrochelidae</b>						
<i>Neopodocinum mrciaki</i> Sellnick, 1958	3	0.55	102	0.65	60	alder (high alt.)
<i>Macrocheles decoloratus</i> (C. L. Koch, 1839)	4	1.30	136	0.87	40	alder (high alt.)
<i>M. montanus</i> Willmann, 1951	4	1.30	136	0.87	40	alder (low and high alt.), fir-beech
<i>Geholaspis longispinosus</i> (Kramer, 1876)	4	0.84	136	0.87	60	alder (low and high alt.)
<b>Pachylaelapidae</b>						
<i>Pachylaelaps furcifer</i> Oudemans, 1903	12	2.30	408	2.60	60	alder (low and high alt.)



<i>P. pectinifer</i> (G & C. Canestrini, 1882)	2	0.55	68	0.43	40	oak-hornbeam
<i>Otopachys vysotskajae</i> Koroleva, 1976	2	0.55	68	0.43	40	alder (low and high alt.)
<i>O. suecicus</i> Sellnick, 1950	7	0.55	238	1.52	100	alder (low and high alt.), beech, oak-hornbeam
<i>Pachyseius humeralis</i> Berlese, 1910	3	0.55	102	0.65	60	fir-beech, oak-hornbeam
<b>Laelapidae</b>						
<i>Pseudolaelaps doderoi</i> (Berlese, 1910)	16	3.03	544	3.46	80	beech, oak-hornbeam
<i>Hypoaspis aculeifer</i> (Caneastrini, 1883)	21	2.59	714	4.55	100	alder (low and high alt.), beech, fir-beech
<i>H. oblonga</i> (Halbert, 1915)	2	0.55	68	0.43	40	alder (low and high alt.)
<i>H. miles</i> Berlese, 1892	4	1.10	136	0.87	40	oak-hornbeam
<b>Eviphididae</b>						
<i>Eviphis ostrinus</i> (C. L. Koch, 1836)	5	1.22	170	1.08	60	alder (low and high alt.)
<b>Zerconidae</b>						
<i>Zercon pinicola</i> Halaskova, 1970	1	0.45	34	0.22	20	alder (low alt.)
<i>Z. triangularis</i> C. L. Koch, 1836	7	1.67	238	1.52	60	alder (low and high alt.)
<i>Z. hungaricus</i> Sellnick, 1958	11	1.48	374	2.38	80	oak-hornbeam
<i>Z. fageticola</i> Halaskova, 1969	2	0.55	68	0.43	40	fir-beech, oak-hornbeam
<i>Prozercon plumatus</i> (Aoki, 1966)	15	4.64	510	3.25	60	alder (low and high alt.)
<i>P. kochi</i> Selnick, 1943	40	7.65	1360	8.66	100	alder (low and high alt.)
<i>P. fimbriatus</i> (C. L. Koch, 1839)	74	2.39	2516	16.02	100	alder (low alt.), beech, fir-beech, oak-hornbeam
<i>P. traegardhi</i> (Halbert, 1923)	31	5.17	1054	6.71	100	alder (low and high alt.), beech, fir-beech, oak-hornbeam
<i>P. sellnicki</i> Halaskova, 1963	2	0.55	68	0.43	40	oak-hornbeam
<i>Prozercon</i> sp.	2	0.89	68	0.43	20	beech
Total	467		15878	100		

The high numbers of subdominant, recedent, subrecedent, accessory, and accidental species (taking into account the dominance and constancy indices) in beech, fir-beech, low-altitude alder, and oak-hornbeam forests indicate an unfavourable influence of environmental conditions (sandy soils with little organic matter, sometimes basic, dry) on stability of the studied populations. Their trophic spectrum determines high mobility, so it is possible that the identified species migrate to the study areas from adjacent ecosystems. Some authors show that predatory mites are influenced by the soil horizon and the period of collecting and not by altitude, without any significant interaction (SADAKA & PONGE 2003). Each of the investigated ecosystems had a different species composition.

The dendrogram (Fig. 1) shows some similarities between species diversity of gamasids, explicable through the similarities in environmental conditions. For example, higher values of Jaccard index for Cumpătu vs. Călugăreni, as well as Balotești vs. Băneasa, were probably due to the same substrate (alluvial soil) and to the same primary producers (alder forests) and at Șotriile vs. Clinceanca due to the close values of soil moisture content and acidity. Low values of the Jaccard index were obtained for Azuga Valley compared to Lunca Mare, Șotriile, Băneasa, and Balotești (Table 4). This was probably due to differences in primary producers, altitude, soil type, and acidity for Lunca Mare vs. Azuga Valley; primary producers, moisture content, altitude, acidity, and soil type for Șotriile vs. Azuga Valley; primary producers and altitude for Băneasa vs. Azuga Valley; and soil moisture content for Balotești vs. Azuga (Fig. 1).

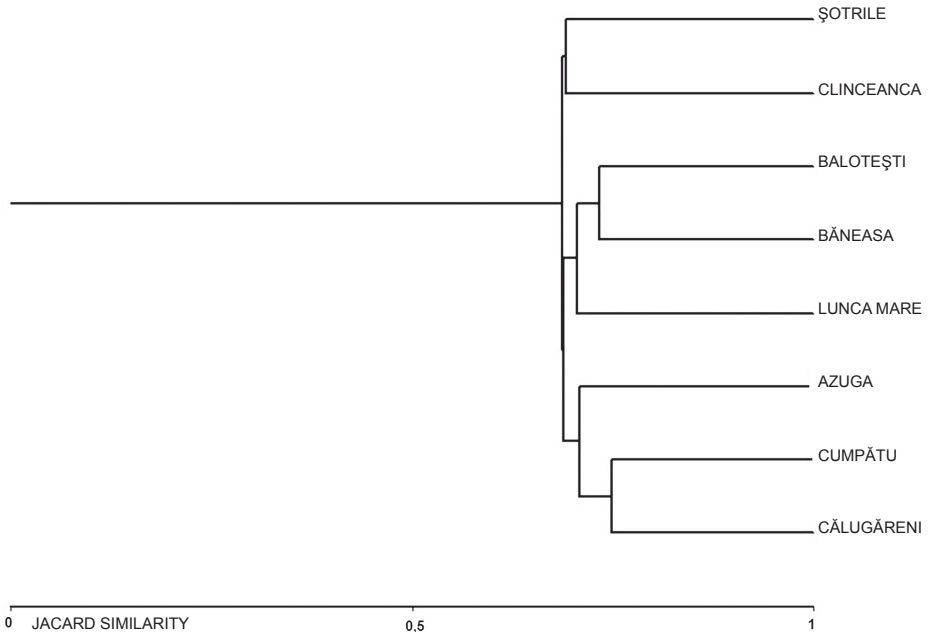


Fig. 1. Dendrogram based on the Jaccard similarity index of species composition between studied forest ecosystems

Table 4. Jaccard index of similarity of gamasid species composition between the studied forest ecosystems. Values &gt; 0.60 are marked in bold

	Clinceanca	Azuga	Cumpătu	Lunca Mare	Șotriile	Băneasa	Balotești
Călugăreni	<b>0.62</b>	0.59	<b>0.75</b>	0.56	0.60	0.54	0.57
Clinceanca		0.47	<b>0.67</b>	<b>0.69</b>	<b>0.69</b>	<b>0.65</b>	0.57
Azuga			<b>0.71</b>	0.33	0.33	0.37	0.38
Cumpătu				0.49	0.52	<b>0.69</b>	<b>0.69</b>
Lunca Mare					<b>0.62</b>	<b>0.71</b>	0.57
Șotriile						0.43	0.55
Băneasa							<b>0.73</b>

## CONCLUSIONS

This study highlights the structural differences in gamasid communities. Considering species diversity and the dominance structure of gamasid mite communities, the alder forests of Călugăreni and oak-hornbeam forest of Băneasa had proper environmental conditions for mite development. The lowest gamasid abundance and species richness were recorded in the fir-beech forest of Șotriile. The small number of identified gamasid species, which usually had a low number of individuals, showed that most of the studied forests were unstable ecosystems.

Geographical position, abiotic factors (microclimate, soil type) and biotic ones (vegetation structure: herbaceous plants, shrubs, trees) determine the variation in species composition in each of the investigated forest ecosystems. High soil moisture content in correlation with vegetation structure caused a high species diversity in Băneasa and Călugăreni forests. By contrast, steep slopes (30-40°) of Lunca Mare and Șotriile as well as the slightly basic soil pH of Azuga and Balotești forests could determine the decreased species diversity of gamasids.

The structure of gamasid communities depended also on the geographical position and abiotic factors, which characterized the studied forest ecosystems. Development of Gamasina communities is influenced by microclimate, depending on vegetation structure (herbs, shrubs or trees), and on the litter and humus layers (the trophic reservoirs).

The dominant identified gamasid species differed in population structure, due to the various types of habitat (including vegetation layers), which offer several kinds of trophic sources. In this way, each studied forest ecosystems was described by the dominant gamasid species, with specific ecological preferences. These species can be considered as bioindicators for each type of studied forest ecosystems.

**Acknowledgements:** This study was funded by project no. RO1567-IBB01/2013 “Researches concerning the relation between biodiversity and functions in some ecosystems from the Romanian Carpathians” from the Institute of Biology, Romanian Academy, Bucharest. Thanks are due to anonymous referees for their useful suggestions, constructive comments and advice.

## REFERENCES

- BEDANO J., RUF A. 2010. Sensitivity of different taxonomic levels of soil Gamasina to land use and anthropogenic disturbances. *Agric. For. Entomol.* 12: 203–212.
- BERG P. M., BENGSSON J. 2007. Temporal and spatial variability in soil food web structure. *Oikos* 116: 1789–1804.
- BOTNARIUC N., VĂDINEANU A. 1982. *Ecologie [Ecology]*. Editura Didactică și Pedagogică, București (in Romanian).
- CHRISTIAN A. 1995. Succession of Gamasina in coal mine areas in Eastern Germany. *Acta Zool. Fenn.* 196: 380–381.
- COLEMAN D. C., WHITMAN W. B. 2005. Linking species richness, biodiversity and ecosystem function in soil systems. *Pedobiologia* 49: 479–497.
- DONIȚĂ C., POPESCU A., PAUCĂ COMĂNESCU M., MIHĂILESCU S., BIRIȘ I. 2005. Habitatele din România [Habitats from Romania]. Editura Tehnică și Silvică, 269–271 (in Romanian).
- ENGELMANN H. D. 1978. Zur Dominanzklassifizierung von Bodenarthropoden [Dominance classification of the epigeal arthropods]. *Pedobiologia* 18: 378–380.
- FALCĂ M., VASILIU-OROMULU L., SANDA V., PAUCĂ-COMĂNESCU M., HONCIUC V., MAICAN S., PURICE D., TATOLE A., STĂNESCU M., ONETE M., BIȚĂ-NICOLAE C., FIERA C., ȘINCU E. D. 2005a. The Biodiversity of the Alder Forest from the Mountainous Valleys of the Girbova Massif. *Proceedings of the Institute of Biology, Romanian Academy* 7: 15–33.
- FALCĂ M., VASILIU-OROMULU L., SANDA V., PAUCĂ-COMĂNESCU M., HONCIUC V., MAICAN S., PURICE D., TATOLE A., STĂNESCU M., ONETE M., BIȚĂ-NICOLAE C., FIERA C., ȘINCU E. D. 2005b. The Study of Preliminary and Secondary Producers from Some Alder Depression Forests of the Plain Zone. *Proceedings of the Institute of Biology, Romanian Academy* 7: 33–49.
- GILYAROV M. S., BREGETOVA N. G. 1977. *Opređeliteli obitayushchikh v pochve kleshchei Mesostigmata [A Key to Soil-Dwelling Mites (Mesostigmata)]*. Akademia Nauk USSR, Zoologicheskii Institut Evolyucionoi Morfologii i Ekologii zhivotnikh im A.H. Savertova, Izd. Nauka, Leningrad.
- GWIAZDOWICZ D., KLEMT I. 2004. Mesostigmatic mites (Acari, Gamasida) in selected microhabitats of the Biebrza National Park (NE Poland). *Biological Lett.* 41: 11–19.
- GWIAZDOWICZ D. 2007. Ascid mites (Acari, Gamasina) from selected forest ecosystems and microhabitats in Poland. University Augusta Cieszkowskiego, Poznań.
- GULVIK M. E. 2007. Mites (Acari) as indicators of soil biodiversity and land use monitoring: a review. *Pol. J. Ecol.* 55: 415–450.
- HYATT K. H. 1980. Mites of the subfamily Parasitinae (Mesostigmata: Parasitidae) in the British Isles. *Bull. Br. Mus., Zoology series* 38: 237–378.
- KARG W. 1993. Acari (Acarina), Milben Parasitifomes (Anactinochaeta) Cohors Gamasina Leach – Raubmilben [Acari (Acarina) supraorder Parasitifomes (Anactinochaeta) Cohors Gamasina Leach – predatory mites]. *Spektrum Akademischer Verlag* (in German).
- KOEHLER H. H. 1997. Mesostigmata (Gamasina, Uropodina) efficient predators in agroecosystems. *Agricult. Ecosys. Environ.* 62: 105–117.
- KOEHLER H. H. 1999. Predatory mite’s (Gamasina, Mesostigmata). *Agricult. Ecosys. Environ.* 74: 395–410.

- LINDBERG N., BENGTSSON J. 2006. Recovery of forest soil fauna diversity and composition after repeated summer droughts. *Oikos* 114: 494–506.
- KOOLJMAN A. M., MOURIK J. M., SCHILDER M. L. M. 2009. The relationship between N mineralization or microbial biomass N with micromorphological properties in beech forest soils with different texture and pH. *Biol. Fertil. Soils* 45: 449–459.
- MANU M. 2008. Structure and dynamics of the predatory mites (Acari: Mesostigmata-Gamasina) from Bucharest. In: *Species Monitoring in the central parks of Bucharest* (CRĂCIUN I., Ed.), pp. 68–78, Editura Ars Docendi, Bucharest University.
- MANU M., HONCIUC V. 2010a. Ecological research on the soil mite's populations (Acari: Mesostigmata-Gamasina, Oribatida) from forest ecosystems near Bucharest city. *Rom. J. Biol. – Zool.* 55: 19–30.
- MANU M., HONCIUC V. 2010b. Cercetări ecologice asupra populațiilor de gamaside (Acari: Mesostigmata) din solurile unor ecosisteme forestiere din Masivul Bucegi-România. [Ecological researches on gamasid populations (Acari: Mesostigmata) from soils of some forest ecosystems from Bucegi Massif- Romania]. Structure and dynamics. *Ars Docendi*, University Bucharest (in Romanian).
- MARAUN M., MARTENS H., MIGGE S., THEENHAUS A., SCHEU S. 2003. Adding to 'the enigma of soil animal diversity': fungal feeders and saprophagous soil invertebrates prefer similar food substrates. *Eur. J. Soil Biol.* 39: 85–95.
- MÁŠAN P. 2003. Macrochelid mite's of Slovakia (Acari, Mesostigmata, Macrochelidae). Institute of Zoology, Bratislava, Slovak Academy of Science.
- MÁŠAN P., FENDA P. 2004. Zerconid mite's of Slovakia (Acari, Mesostigmata, Zerconidae). Institute of Zoology, Bratislava, Slovak Academy of Science.
- MÁŠAN P. 2007. A review of the family Pachylaelapidae in Slovakia with systematics and ecology of European species (Acari: Mesostigmata: Eviphidoidea). Institute of Zoology, Slovak Academy of Science (Bratislava).
- MÁŠAN P., HALLIDAY B. 2010. Review of the European genera of Eviphididae (Acari: Mesostigmata) and the species occurring in Slovakia. *Zootaxa* 2585: 1–122.
- METZ L. J. 1971. Vertical movement of the Acarina under moisture gradients. *Pedobiologia* 11: 262–268.
- MOORE J. C., MCCANN K., RUITER P. C. 2005. Modeling trophic pathways, nutrient cycling, and dynamic stability in soils. *Pedobiologia* 49: 499–510.
- MORAZA M. 2006. Efecto de la degradación de un encinar de *Quercus rotundifolia* en la comunidad de ácaros cryptostigmados y mesostigmados (Acari: Cryptostigmata, Mesostigamata) [Degradation effects of forest with *Quercus rotundifolia* on cryptostigmatid and mesostigmatid mite communities (Acari: Cryptostigmata, Mesostigamata)]. *Revista Ibérica de Aracnología* 13: 171–182 (in Spanish).
- ONETE M., PAUCĂ-COMĂNESCU M. 2008. Heavy metal content assessment in plants from Bucharest. In: *Species Monitoring in the central parks of Bucharest* (CRĂCIUN I., Ed.), pp. 23–50, *Ars Docendi*, University Bucharest.
- SADAKA N., PONGE J. F. 2003. Soil animal communities in holm oak forests: influence of horizon, altitude and year. *Eur. J. Soil Biol.* 39: 197–207.
- SALMANE I. 2000. Investigation of the seasonal dynamics of soil Gamasina mites (Acari: Mesostigmata) in *Pinaceum myrtilosum*. *Latvia. Ekologia* 19: 245–252.
- SALMANE I., BRUMELIS G. 2010. Species list and habitat preference of mesostigmata mites (Acari, Parasitiformes) in Latvia. *Acarologia* 50: 373–394.
- SELVIN S., VACCA A. 2004. *Biostatistics. How it works*. Pearson Education, UK.
- SKORUPSKI M. 2001. Mites (*Acari*) from the order Gamasida in the Wielkopolski National Park. *Fragm. Faun.* 44: 129–167.
- SKORUPSKI M., BELTER W., KAMCZYC J., WIERZBICKA A. 2008. Soil mites (Acari, Mesostigmata) of the 'Torfowiska Doliny Izery' Reserve in the Sudety Mountains. *Soil Organisms* 80: 261–270.

- SKORUPSKI M., BUTKIEWICZ G., WIERZBICKA A. 2009. The first reaction of soil mite fauna (Acari, Mesostigmata) caused by conversion of Norway spruce stand in the Szklarska Poręba Forest District. *J. Forest Sci.* 55: 234–243.
- WALTER D. E., PROCTOR H. 1999. *Mites – Ecology, Evolution and Behaviour*. UNSW Press, Everbest Printing Hong-Kong.